



# Exposure to a Mixture of Metals and Growth Indicators in 6–11-Year-Old Children from the 2013–2016 NHANES

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## Abstract

Lead (Pb), mercury (Hg), and fluoride (F) exposure during childhood is of concern owing to their toxicity. Also, evidence suggests that high and low exposure levels to manganese (Mn) and selenium (Se) during this vulnerable period are associated with an increased risk of adverse health effects. A reduced growth is associated with high Pb and F exposure; however, little is known about their impact on children's body size, and there is a lack of consensus on the effects of Hg, Mn, and Se exposure on children's anthropometric measures. This is particularly true for childhood metal co-exposures at levels relevant to the general population. We investigated the joint effects of exposure to a metal mixture (Pb, Mn, Hg, and Se in blood and F in plasma) on 6–11-year-old US children's anthropometry ( $n = 1634$ ). Median F, Pb, Mn, Hg, and Se concentrations were 0.3  $\mu\text{mol/L}$ , 0.5  $\mu\text{g/dL}$ , 10.2  $\mu\text{g/L}$ , 0.3  $\mu\text{g/L}$ , and 178.0  $\mu\text{g/L}$ , respectively. The joint effects of the five metals were modeled using Bayesian kernel machine and linear regressions. Pb and Mn showed opposite directions of associations with all outcome measured, where Pb was inversely associated with anthropometry. For body mass index and waist circumference, the effect estimates for Pb and Mn appeared stronger at high and low concentrations of the other metals of the mixture, respectively. Our findings suggest that metal co-exposures may influence children's body mass and linear growth indicators, and that such relations may differ by the exposure levels of the components of the metal mixture.

**Keywords** NHANES · Childhood exposure · Metal mixture · Growth · Body size

## Introduction

Heavy metals pose a threat to human health because they are non-biodegradable and can be deposited in body tissues or organs to produce harm after initial exposure (Al Osman et al. 2019). Children are especially vulnerable because of their rapid growth and still developing detoxification

mechanisms (Rodríguez-Barranco et al. 2013). Indicators of growth and body size in school-age children (e.g., height, weight, and waist circumference) play a role in school achievement (Crooks 1995), and physical and emotional health in childhood and adolescence (Gelander 2006). In addition to lifestyle factors (e.g., diet or physical activity), there has been interest in understanding the effect of environmental exposures on children's anthropometric measures.

A number of studies have examined whether exposure to trace elements (hereafter, “metals”) affect children's body size and growth, but the effects at exposure levels relevant to the general population and co-exposures to metal mixtures remain poorly understood. Generally, the primary sources of metal exposure include ingestion through drinking water or diet, inhalation, and dermal contact (Buckley et al. 2020). Existing studies have investigated prenatal exposures and their effects on body size in utero, at birth or in early life, as well as postnatal exposure on growth throughout childhood. For postnatal exposures, lead (Pb) has received the most attention, with consistent evidence that higher Pb exposure associates with shorter stature (Burns et al. 2017; Deierlein

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et al. 2019; Kerr et al. 2019; Raihan et al. 2018; Yang et al. 2013; Zeng et al. 2019) and may reduce body weight indicators across a wide range of ages (Cassidy-Bushrow et al. 2016; Deierlein et al. 2019; Kerr et al. 2019; Scinicariello et al. 2013; Zeng et al. 2019). Fewer studies have been done on mercury (Hg), with two studies suggesting positive associations with Body Mass Index (BMI) in 0–10-year-old children (Benefice et al. 2008; Gao et al. 2018). In contrast, another study found no association with Hg among children without permanent housing (Fábelová et al. 2018). High fluoride (F) exposure appears to negatively affect children's height (Wang et al. 2007). Manganese (Mn) and Selenium (Se) are essential nutrients and as such are necessary for proper growth and development (Lewicka et al. 2017). Available evidence suggests that Mn and Se follow a U-shape dose–response curve with adverse effects at low and high exposure levels on birth outcomes and BMI among adults and young children (Ortega et al. 2013; Rayman 2012; Signes-Pastor et al. 2019; Wang et al. 2016). To date, nearly all studies considered exposures to a single metal at a time.

Exposure to toxic and nutrient metals occur simultaneously as a mixture in real-life scenarios (Valeri et al. 2017); however, there has been little investigation of metal co-exposure effects on body size and growth during childhood. Three prior studies on postnatal exposures and children's growth examined more than two metals (Fábelová et al. 2018; Gardner et al. 2013; Zeng et al. 2019). Measurement of prenatal exposure to multiple metals has been more common (Bloom et al. 2015; Freire et al. 2019; Gleason et al. 2016; Govarts et al. 2016; Sabra et al. 2017; Thomas et al. 2015); yet none of these studies assessed the effects of metal mixtures.

To better understand the potential risks associated with metal co-exposures at levels relevant to the general population and child anthropometry, we investigated the associations between blood concentration of metals, of nutritional (i.e., Mn and Se) and toxicological interest (i.e., F, Hg, and Pb), on body size and growth measurements of 6–11-year-old US children included in the 2013–2016 National Health and Nutrition Examination Survey (NHANES).

## Materials and Methods

### Study Design and Population

Participants were part of the NHANES, an ongoing cross-sectional survey conducted by the National Center for Health Statistics at the Centers for Disease Control and Prevention (CDC). The NHANES aims to assess the health and nutritional status of the civilian, non-institutionalized population in the US. Participants are selected into NHANES using a stratified, multistage probability sampling strategy based on

selection of counties, blocks, households, and individuals in the households. Details of recruitment and design are available on the NHANES website (NHANES 2016a). Procedures for the NHANES have been approved by the National Center for Health Statistics Research Ethics Review Board. Our current analyses included 6–11-year-old children participating in the 2013–2014 and 2015–2016 NHANES cycles.

### Whole Blood Lead, Manganese, Total Mercury, and Selenium

Venous whole blood samples were collected by phlebotomists; Pb, Mn, Hg, and Se levels were measured using an Inductively Coupled Plasma Mass Spectrometer with Dynamic Reaction Cell Technology (ELAN® DRC II) at the National Center for Environmental Health. Detailed methods of these procedures are published elsewhere (NHANES 2013a, 2015a). The lower limit of detection (LOD) was 0.07 µg/dL for Pb, 0.99 µg/L for Mn, 0.28 µg/L for Hg, and 24.48 µg/L for Se for the 2013–2016 study cycles (NHANES 2013a, 2015a, 2016b, 2018). There were no observations below the LOD for Mn and Se. For Pb, only one observation was below the LOD. For Hg, 639 (39%) observations were below the LOD and they were evenly distributed by sex. Values below the LOD were entered as the LOD divided by the square root of two and included in statistical analyses (Lubin et al. 2004; Schisterman et al. 2006).

### Plasma Fluoride

Concentrations of F from plasma samples were measured at the College of Dental Medicine, Georgia Regents University, Augusta, GA, using an ion-specific electrode. The LOD for this method is ~ 1 µmol/L (0.019 mg/L), which is usually higher than plasma F concentrations. To overcome this, the hexamethyldisiloxane facilitated diffusion method is applied, to transfer F from the plasma samples into an alkaline trapping solution of smaller volume. Details of these methods have been published previously (NHANES 2013b, 2015b). There were no plasma F concentrations below the LOD.

### Body Measures

Body measures of interest, such as weight (kg), waist circumference (cm), upper arm length (cm), and standing height (cm) were assessed in the Mobile Examination Center by health technicians. Children's BMI (kg/m<sup>2</sup>) was calculated from weight and height measurements. Arm length was typically measured on the right side of the body, unless participants had a medical condition, amputation, or a cast on the right side. Weight was measured using a digital scale in kg. Detailed protocols have been previously published (NHANES 2013c, 2016a).

## Covariates

Demographic details for the children, such as age, sex, and race were provided by a proxy in an interview. Socio-economic status was measured as poverty-to-income ratio (PIR). Total caloric intake (kcal) was calculated using a single 24-h dietary recall. An interviewer-administered questionnaire asked about the number of people in the household who smoked tobacco. Smokers in the household (i.e., household members smokers  $\geq 1$ ) was used as a proxy for secondhand smoke. Physical activity was measured through an interviewer-administered questionnaire at the participants' home using the Computer Assisted Personal Interview system. Children reported whether or not they were engaged in physical activities outside of school in the preceding 7 days. They were asked about practicing common sports such as basketball, baseball, bike riding, dancing, etc. A total of thirty-three activities were queried. Then, similar questions were asked about 22 activities at school. For the current analysis, participants received a score of one for each activity they reported and zero otherwise. Two separate scores were calculated by summing the positive responses for the outside-of-school and school-based physical activities.

## Statistical Analyses

For all statistical analyses, we excluded participants with missing values for the covariates. The Spearman's coefficients ( $\rho$ ) were calculated for each metal pair and for anthropometric measurements. BMI, standing height, waist circumference, and upper arm length were approximately normally distributed. Plasma F and blood Pb, Mn, Hg, and Se concentrations were positively skewed. Covariate selection was based on prior studies as well as directed acyclic graphs using the DAGitty software; total calorie intake (kcal, continuous), race (Mexican American, non-Hispanic white, non-Hispanic black or another race, categorical), PIR ( $\leq 1.49$  (the median) vs.  $> 1.49$ ), children's age (years, continuous), smoker/s in the household (yes vs. no), children's sex (boys vs. girls), and outside-of-school and school-based physical activity scores (continuous) were included as potential confounders. The two physical activity scores were uncorrelated (Spearman's  $\rho = 0.10$ ).

Bayesian kernel machine regression (BKMR) was performed primarily as an exploratory method to investigate interactions and joint effects of the five metals using the R package "bkmr" (Bobb et al. 2015, 2018). BKMR models were applied as  $Y_i = h(F_i, Mn_i, Pb_i, Se_i, Hg_i) + \beta^T Z_i + e_i$ , where  $Y$  is the continuous outcome of interest (i.e., BMI, standing height, waist circumference or upper arm length);  $h()$  is an exposure–response function that accommodates nonlinearity and interactions among mixture components: F, Pb, Mn, Hg, and Se concentrations natural

log-transformed, centered and scaled;  $Z$  are the selected covariates and  $\beta$  are the corresponding regression coefficients. All models included 10,000 Markov chain Monte Carlo iterations using the Gaussian kernel, with 5000 used as burn-in. The BKMR model is not able to accommodate sample weights yet, and thus we used unweighted estimation; however, all our models included several covariates that are used to calculate the weights (e.g., age, sex, or race) (Kim et al. 2017; Zhang et al. 2019).

Linear regression analyses were then conducted to further evaluate and quantify the associations between metal exposure and children's anthropometric measures. The F, Pb, Mn, Hg, and Se concentrations were log-transformed to approximate normal distribution and divided by their interquartile range (IQR) for normalization before inclusion in the regression models. Anthropometric measurements (i.e., BMI, standing height, waist circumference, and upper arm length) were entered in the models as dependent variables. A threshold of  $\alpha = 0.05$  was used to define associations as statistically significant. We first conducted analyses in the overall sample (adjusted for sex), then stratified by sex. All analyses and graphics were performed using R version 3.5.1 (R Core Team 2014).

## Results

### Study Population Characteristics

Out of 2703 children included in the 2013–2016 NHANES cycles, 2092 had blood samples analyzed for F, Pb, Mn, Hg and Se concentrations. Of those, 2049 had complete information on body size and growth parameters (i.e., BMI, standing height, waist circumference, and upper arm length). Finally, 1634 children, evenly distributed across cycles ( $n = 835$  and  $n = 799$  in 2013–2014 and 2015–2016, respectively) and sex ( $n = 839$  boys and  $n = 795$  girls), had complete data on the covariates and were included in the statistical analyses (Figure S1). The biochemical, socio-economic, and anthropometric characteristics of the children included in the analysis ( $n = 1634$ ) did not differ from those who were excluded ( $n = 1069$ , Table S1). The studied children had a median age of 9 years and median F, Pb, Mn, Hg, and Se concentrations were  $0.3 \mu\text{mol/L}$ ,  $0.5 \mu\text{g/dL}$ ,  $10.2 \mu\text{g/L}$ ,  $0.3 \mu\text{g/L}$ , and  $178.0 \mu\text{g/L}$ , respectively. The baseline study population characteristics did not differ by sex (Table 1). Blood metals concentrations were weakly correlated and had a Spearman's  $\rho$  ranging from  $-0.12$  to  $0.12$  (Figure S2). On the other hand, children's body size indicators were highly correlated with a Spearman's  $\rho > 0.9$  for BMI and waist circumference (Figure S3).

**Table 1** Selected characteristics of 6–11-year-old children from the 2013–2016 NHANES

Characteristics	Overall sample ( <i>n</i> = 1634)	Boys ( <i>n</i> = 839)	Girls ( <i>n</i> = 795)
Age (years)	9.0 (7.0–10.0) <sup>1</sup>	8.0 (7.0–10.0)	9.0 (7.0–10.0)
BMI (kg/m <sup>2</sup> )	17.7 (15.7–21.3)	17.3 (15.6–21.1)	18.2 (15.8–21.3)
Standing height (cm)	134.3 (125.8–144.2)	134.3 (126.1–143.4)	134.5 (125.3–145.0)
Waist circumference (cm)	62.2 (56.3–72.6)	61.2 (55.9–72.0)	64.1 (56.9–73.6)
Upper arm length (cm)	28.6 (26.3–32.0)	28.6 (26.3–0.8)	28.8 (26.2–31.3)
Energy (kcal)	1800 (1406–1905)	1858 (1480–2350)	1753 (1362–2168)
Outside of school physical activity score	3.0 (1.0–4.0)	3.0 (1.0–4.0)	2.0 (1.0–4.0)
At-school physical activity score	0.0 (0.0–1.0)	0.0 (0.0–1.0)	0.0 (0.0–1.0)
Poverty-to-income ratio			
≤ median (1.49)	855 (52) <sup>2</sup>	431 (51)	424 (53)
> median (1.49)	779 (48)	408 (49)	371 (47)
Race			
Mexican American	571 (35)	275 (33)	296 (37)
Non-Hispanic white	450 (28)	242 (24)	208 (26)
Non-Hispanic black	398 (24)	211 (25)	187 (23)
Another race	215 (13)	111 (13)	104 (13)
Household members that smoke	431 (26)	233 (28)	198 (25)
Cycles			
2013–2014	835 (51)	436 (52)	399 (50)
2015–2016	799 (49)	403 (48)	396 (50)
Metal/trace element concentrations			
F <sup>4</sup>	0.3 (0.1, 0.3–0.5, 4.0) <sup>3</sup>	0.3 (0.1, 0.3–0.5, 3.9)	0.3 (0.1, 0.2–0.4, 4.0)
Pb <sup>5</sup>	0.5 (0.1, 0.4–0.8, 5.8)	0.5 (0.1, 0.4–0.8, 5.0)	0.5 (0.1, 0.4–0.7, 5.8)
Mn <sup>6</sup>	10.2 (3.8, 8.4–12.6, 24.3)	9.8 (3.8, 7.9–12.2, 23.4)	10.7 (4.6, 8.8–13.0, 24.3)
Hg <sup>6</sup>	0.3 (0.2, 0.2–0.5, 10.0)	0.3 (0.2, 0.2–0.5, 7.6)	0.3 (0.2, 0.2–0.5, 10.0)
Se <sup>6</sup>	178.0 (110.2, 167.7–190.2, 238.5)	177.6 (110.2, 167.1–189.5, 238.5)	178.5 (138.7, 168.6–190.5, 227.0)

*BMI* = Body Mass Index

<sup>1</sup>Continuous variables = median (IQR)

<sup>2</sup>Categorical variables = *n* (%)

<sup>3</sup>Metal/trace element concentrations = median (min, IQR, max)

<sup>4</sup>Measured in plasma (μmol/L)

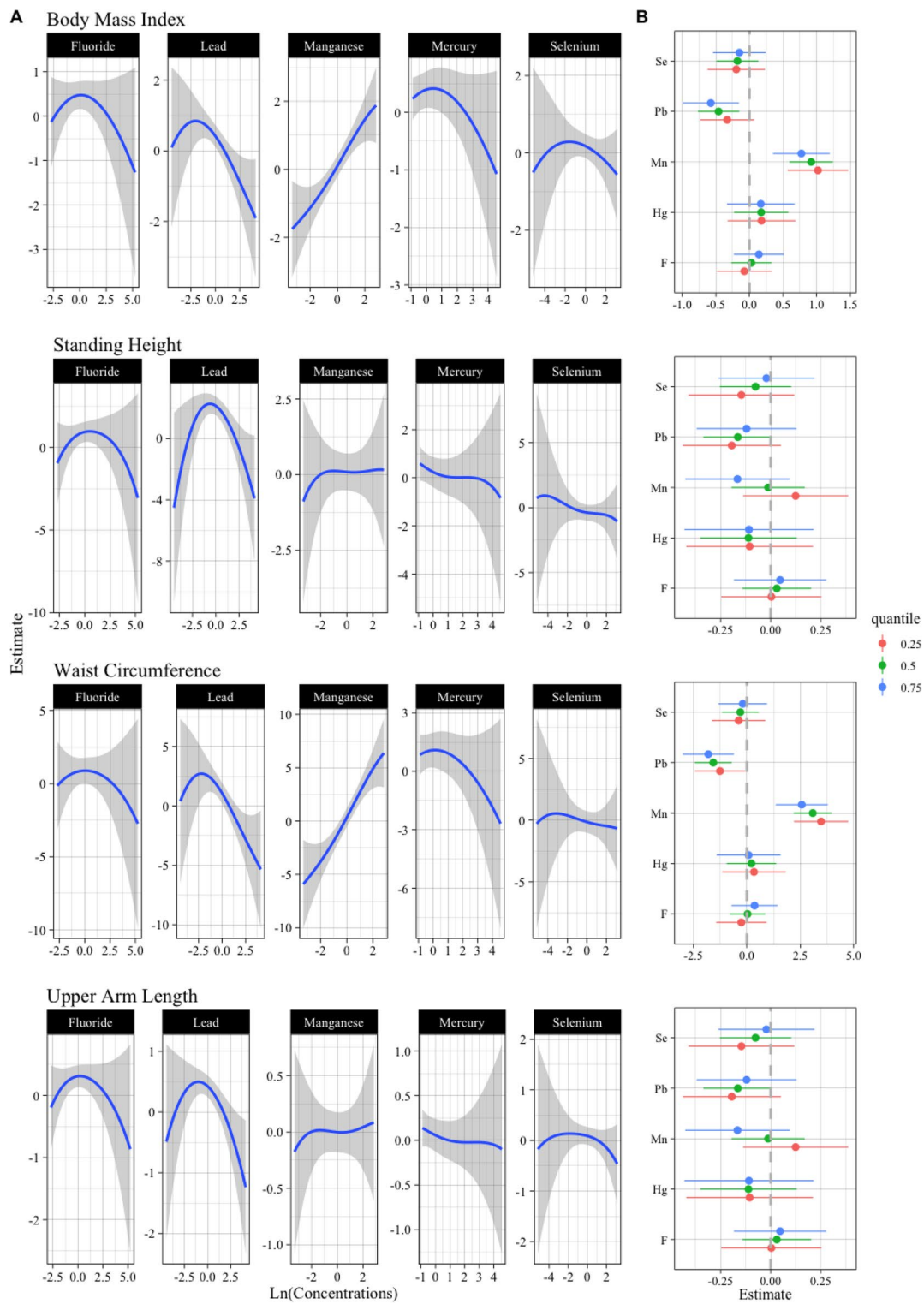
<sup>5</sup>Measured in blood (μg/dL)

<sup>6</sup>Measured in blood (μg/L)

## Bayesian Kernel Machine Regression

The main findings from the BKMR and linear regression models overall and stratified by sex are summarized in Table S2. Mn was positively associated with children's BMI and waist circumference (Fig. 1A), and this association appeared stronger at lower percentiles of the other metals of the mixture (Fig. 1B). Conversely, Pb was negatively associated with BMI and waist circumference (Fig. 1A), and this association appeared stronger at higher percentiles of the other metals of the mixture (Fig. 1B). These relationships appeared linear; however, there was variability at lower Pb concentrations (Fig. 1A). Overall, higher variability at low and high metal concentrations is expected to be related to

the limited number of observations at such exposure levels. When we added variable selection in the BKMR models, Mn and Pb were selected for inclusion in more than 50% of iterations [Posterior Inclusion Probability (PIP) > 0.5] (Figure S4), and thus they were deemed to be important contributors of the variability of the outcomes (Barbieri and Berger 2004; Coker et al. 2018; Laue et al. 2020). Effect estimate changes for an IQR increase of each component across the 25th, 50th, and 75th percentiles of the other metals of the mixture suggested Mn and Pb interactions (Fig. 1B). The bivariate exposure–response functions support the findings of a Mn–Pb interaction with BMI and waist circumference (Figure S5). Generally, we found similar results in the sex-stratified analyses (Figures S6 and S7); however,



**Fig. 1** Metal concentrations and anthropometric indicators of 6–11-year-old children from the 2013 to 2016 NHANES. BKMR dose–response function and interactions within components of the metal mixture ( $n=1634$ ). Models are adjusted for total calorie intake (kcal, continuous), race (Mexican American, non-Hispanic white, non-Hispanic black or another race, categorical), poverty-to-income ratio ( $\leq 1.49$  (the median) vs.  $> 1.49$ ), smoker/s in the household (yes vs. no), children’s age (years, continuous), outside-of-school and at-

school activity scores (continuous), and child’s sex (boys vs. girls). **A** Univariate exposure–response functions and 95% confidence bands for each component with the other components fixed at the median. **B** Single component association (estimates and 95% credible intervals, gray dashed line at the null). This plot compares children’s size when a single component is at 75th vs. 25th percentile, when all the other exposures are fixed at either the 25th, 50th, or 75th percentile

with less statistical precision and appearance of nonlinear dose–response, especially among boys. No clear associations were observed with the remaining components of the metal mixture and children’s anthropometric measurements (Fig. 1B).

### Multiple Metal Linear Regression

Using multiple linear regression analyses, an IQR difference in blood Mn was associated with a 0.88 kg/m<sup>2</sup> (95% Confidence Interval (CI) 0.59, 1.18) and 2.46 cm (95% CI 1.69, 3.25) higher BMI and waist circumference, respectively (Table 2). In addition, an IQR change in blood Mn was associated with a 0.20 cm (95% CI 0.05, 0.35) and 0.70 cm (95% CI 0.21, 1.19) difference in upper arm length and standing height, respectively (Table 2). The standing height estimate appeared stronger among boys (Table S3). Conversely, higher blood Pb was associated with a lower BMI [−0.47 cm (95% CI −0.72, −0.21)] and waist circumference [−1.29 cm (95% CI −1.97, −0.61)], as well as lower upper arm length [−0.24 cm (95% CI −0.36, −0.11)] and standing height [−0.70 cm (95% CI −1.13, −0.27)] (Table 2). The positive and negative associations between Mn and Pb concentrations with children’s anthropometry were each consistent in our sex-stratified analyses. We noticed a borderline significant two-way interaction between Mn and Pb with BMI (*p*-value = 0.057) but attenuated with waist circumference (*p*-value = 0.133). Further, we observed that an IQR difference in blood Hg concentration was associated with a 0.50 cm (95% CI −0.97, −0.02) lower standing height

(Table 2), particularly among girls [−0.67 (95% CI −1.34, 0.00)]. Also, among girls, a negative association was found between blood Hg and upper arm length but with wide confidence intervals (Table S3). As in the BKMR analysis, we did not identify associations with other components of the mixture. Additionally, we performed regression models adding metal quantile concentrations (categorical) as independent variable, and the results supported the linear associations previously described (Table S4).

### Discussion

Our findings, based on a population of US children with a median age of 9 years participating in the NHANES 2013–2016 cycles, suggest that exposure to a metal mixture of F, Mn, Pb, Se, and Hg may influence child body size and growth depending on the exposure levels measured in blood. Plasma F, and whole blood Mn, Pb, Se, and Hg concentrations are considered an accurate/valid biomarker of exposure (Buckley et al. 2020), and the levels found in our population are comparable to other population-based studies (Bose-O’Reilly et al. 2010; Grandjean 2019; Henn et al. 2010; Hrubá et al. 2012; Kobayashi et al. 2019; Leite et al. 2015; Liu et al. 2014). We observed that Pb concentrations were related to lower BMI and waist circumference at higher concentrations of the other metals of the mixture with little evidence of nonlinearity. A negative association was also observed between blood Pb and children’s upper arm length and standing height, especially among boys.

**Table 2** Difference in the BMI, standing height, weight circumference and upper arm length for each IQR difference in metal concentration of 6–11-year-old children from the 2013–2016 NHANES

	Total sample ( <i>n</i> = 1634)			
	BMI (kg/m <sup>2</sup> )	Standing height (cm)	Waist circumference (cm)	Upper arm length (cm)
Multiple elements $\beta$ (95% CI) <sup>a</sup>				
F	0.00 (−0.24, 0.25)	0.00 (−0.42, 0.41)	0.10 (−0.56, 0.76)	−0.04 (−0.16, 0.09)
Pb	<b>−0.47</b> ( <b>−0.72, −0.21</b> )	<b>−0.70</b> ( <b>−1.13, −0.27</b> )	<b>−1.29</b> ( <b>−1.97, −0.61</b> )	<b>−0.24</b> ( <b>−0.36, −0.11</b> )
Mn	<b>0.88</b> ( <b>0.59, 1.18</b> )	<b>0.70</b> ( <b>0.21, 1.19</b> )	<b>2.46</b> ( <b>1.69, 3.25</b> )	<b>0.20</b> ( <b>0.05, 0.35</b> )
Hg	−0.05 (−0.33, 0.23)	<b>−0.50</b> ( <b>−0.97, −0.02</b> )	−0.20 (−0.95, 0.55)	−0.13 (−0.27, 0.01)
Se	−0.17 (−0.44, 0.09)	−0.20 (−0.65, 0.25)	−0.31 (−1.01, 0.40)	−0.07 (−0.20, 0.06)

Statistically significant results are highlighted in bold

<sup>a</sup>Based on linear models adjusted for total calorie intake (kcal, continuous), race (Mexican American, non-Hispanic white, non-Hispanic black or another race, categorical), poverty-to-income ratio ( $\leq 1.49$  (the median) vs.  $> 1.49$ ), children’s age (years, continuous), smoker/s in the household (yes vs. no), outside-of-school and at-school activity scores (continuous), and children’s sex (boys vs. girls). The models include all five metals (F, Pb, Mn, Hg, and Se)

Blood Mn, most likely reflecting its function as an essential nutrient, was related to higher BMI and waist circumference at lower concentrations of the other metals, and the shape of the dose–response curve appeared to be linear. A positive association was also observed between Mn concentrations and upper arm length and standing height, especially among boys. Blood Hg concentrations were related to reduced standing height, and this association appeared stronger among girls.

Lead is a known toxic metal. Indeed, no levels of blood Pb are considered safe and childhood exposures have been associated with a broad spectrum of deleterious health effects (Burns et al. 2017). Thus, the CDC reduced the blood Pb reference value from 10 to 5  $\mu\text{g}/\text{dL}$  in 2012 to minimize risks among vulnerable population groups. Yet, detectable levels of blood Pb among US children persist (Bellinger et al. 2017). In children, high blood Pb concentrations have been associated with lower osteocalcin, a biomarker of bone formation. Pb may interfere with bone cell function, metabolism, and bone mineralization (Mushak et al. 1989; Pounds et al. 1991). For example, Pb may alter circulating levels of hormones required for bone development and maintenance (e.g., 1,25-dihydroxyvitamin D3), as well as the ability of bone cells to respond to hormonal regulation, leading to impaired bone formation. Further, exposure to Pb may have endocrine-disrupting capabilities by reducing responses to hormones that are necessary for growth, such as insulin-like growth factor, and inhibiting the hypothalamic-pituitary-growth axis (Berry et al. 2002; Burns et al. 2017; Deierlein et al. 2019; Fleisch et al. 2013).

Limited epidemiological evidence is available on the effects of low Pb exposure on child growth, particularly as a component of common real-life metal co-exposures. A prior study found that peripubertal blood Pb concentrations in a single measurement were associated with shorter height through age 18 years (Burns et al. 2017). Our study agrees with the prior findings and suggests an inverse association between blood Pb and standing height and upper arm length. In our BKMR models, there was little evidence that Pb interacted with the other components of the metal mixture on these growth measures.

Associations of Pb exposure with indirect estimations of adiposity such as BMI or waist circumference are less commonly reported, and show null (Ballew et al. 1999; Min et al. 2008) and adverse effects (Burns et al. 2017; Cassidy-Bushrow et al. 2016; Deierlein et al. 2019; Little et al. 2009; Schwartz et al. 1986; Scinicariello et al. 2013). Prior studies using NHANES 1976–1980 and 1999–2006 data found higher blood Pb associated with lower BMI (Scinicariello et al. 2013) and weight (Schwartz et al. 1986) in children. In contrast, a study using NHANES 1988–1994 data did not find associations between blood Pb and children's weight or BMI (Ballew et al. 1999). In our analysis, a consistent

inverse association was observed between Pb exposure and children's BMI and waist circumference that appeared stronger at higher concentrations of the other metals of the mixture.

Children's exposure to Mn is a public health concern that points to complexities in establishing exposure thresholds because of its dual role as an essential nutrient required to maintain health, while it is neurotoxic at high levels (Chung et al. 2015; Signes-Pastor et al. 2018). Evidence suggests that Mn follows a U-shaped dose–response curve on children's neurodevelopment (Henn et al. 2010) and birth size (Chen et al. 2014; Guan et al. 2014; Signes-Pastor et al. 2019; Xia et al. 2016). However, optimal Mn levels have not been defined yet, and there are scant studies on Mn exposure and children's body size and growth. In this study, we found a linear positive association with children's body size, with evidence of antagonistic effects of Mn and Pb exposure on children's BMI and waist circumference. Specifically, the positive association between BMI and waist circumference and blood Mn levels appeared stronger at a lower level of Pb and other mixture components. We found that upper arm length was higher with higher Mn levels, and higher standing height among boys. The Mn exposure levels assessed in our study population appeared to be beneficial for 6–11-year-old children's growth.

Little is known about the effects of Hg exposure on physical growth. However, it has been suggested that in growing children a mild exposure to Hg associated with a fish-based diet consistent with health recommendations may overcome the known adverse risks of Hg exposure (Benefice et al. 2008; Bose-O'Reilly et al. 2010; Gao et al. 2018; Stratakis et al. 2020). Ingestion of polyunsaturated fatty acids, essential fatty acids with important physiological active functions, are among the main nutritional advantages of a fish-based diet (Benefice et al. 2008; Stratakis et al. 2020). However, there is still controversy regarding the benefits or disadvantages of fish consumption (Mozaffarian and Rimm 2006; Stratakis et al. 2020). A cross-sectional study from China reported a positive association between blood Hg and children's anthropometry by comparing 0–6-year-old children from suburban and rural areas with median blood Hg concentrations of 1.34  $\mu\text{g}/\text{L}$  and 1.09  $\mu\text{g}/\text{L}$ , respectively (Gao et al. 2018). Contrary, we observed that standing height was 0.5 cm lower in children for each IQR difference in Hg concentrations and that upper arm length was about 0.1 cm lower, but the latter association did not reach statistical significance. The Hg effect estimates did not appear to be altered by the other components of the metal mixture. Nevertheless, these findings should be interpreted cautiously given the high proportion of Hg concentrations below the LOD (i.e., 39%). The median blood Hg in our study population was about fourfold lower compared to the levels reported in the study from China (Gao et al. 2018), which suggests

a low seafood consumption. Indeed, low fish and shellfish intake has been previously reported among US children of 6–11 years from NHANES 2013–2016 cycles: only 5.8% reported seafood consumption at least twice per week (Terry et al. 2018). Thus, intake of fish and shellfish in our study population is not expected to affect anthropometric indicators, and thus we did not adjust our core models for consumption of marine products. Similarly, sensitivity analysis adjusting linear regression models for marine products including fish and shellfish intake in the past 30 days showed consistent results (data not shown).

Selenium is an essential nutrient, and is a cofactor required by a number of enzymes with antioxidant functions. Its deficiency may lead to modifications in cellular antioxidative capacity and the appearance of a number of diseases. Children are at particular risk for Se deficiency since their rapid growth includes a high demand for Se (Ortega et al. 2013). For instance, lower serum Se levels have been associated with excess weight and reduced height (Lander et al. 2015; Ortega et al. 2013), which supports the epidemiologic data that children need a small amount of Se for normal growth and development (Fábelová et al. 2018). However, our findings provide little evidence of an association between blood Se and children's anthropometric indicators as no evidence of Se deficiency was observed from their Se blood concentrations. Evidence of Se interactions with the other components of the metal mixture is also limited.

Most of the literature on F exposure is focused on cognitive effects, and suggests a dose–response dependent neurotoxicity at high concentrations but also possibly at even levels below the currently accepted 0.7 mg/L in drinking water for preventive dentistry purposes in the US (Grandjean 2019). Nonetheless, a study from China investigated the effects of F exposure from drinking water on 6–12-year-old children's growth, and reported a decreased height among children highly exposed to F at 8.3 mg/L compared to those consuming water at a F level of 0.5 mg/L (Wang et al. 2007). Also, among adults, an increase in bone fractures was associated with drinking water consumption that contained 4 mg/L of F [Institute of Medicine (U.S.) 1997]. There is also evidence that F is accumulated in bone and reduces calcium uptake at high F exposure levels, thereby influencing children's growth and bone strength [Institute of Medicine (U.S.) 1997; Wang et al. 2007]. However, in our population of US school children with relatively low plasma F levels, we did not identify any associations between F and anthropometric measurements.

Drinking water is expected to play a major role in metal exposure for our study population in addition to diet and other environmental factors (Buckley et al. 2020; Signes-Pastor et al. 2018). Several metals are naturally occurring in public drinking water sources and others are added for safety and health benefits (e.g., F to prevent cavities), migrated

from water pipes (e.g., Pb) or related to industrial activities (Chowdhury et al. 2016; Grandjean 2019; Ljung et al. 2011). Leafy vegetables, nuts, grains, and animal products are good sources of Mn and Se, while the consumption of certain seafood products is associated with ingestion of Hg (Lucchini et al. 2017; Navarro-Alarcon and Cabrera-Vique 2008; Stratakis et al. 2020). Nevertheless, metal content in foodstuff can be exacerbated when grown in contaminated soils (Carbonell-Barrachina et al. 2009; Kachenko and Singh 2006). Contact with chips and dust from toys and household Pb-based paint is of particular concern as a Pb exposure source among children (O'Connor et al. 2018; Shen et al. 2018).

In interpreting our findings, it is important to consider some limitations. First is the cross-sectional nature of the data, where exposures and anthropometric indicators were measured at the same time point. Thus, we cannot draw conclusions regarding temporality of the relationship between metal exposure and children's body size and growth; these associations will need to be replicated using prospective data with serial measurements in the future. Second, a single blood sample was used to assess metal exposure. Metal concentrations may fluctuate over time, and the use of a single blood sample to assess exposure is based on the assumption that the single measure represents usual exposure levels. Third, our study findings may not be generalizable beyond the age range of 6–11 years because children continue to grow until about 20 years of age, and different toxicants/nutrients may have different critical exposure windows based on the outcomes of interest. In contrast, our study is among the first to flexibly model the association between both nutrient and toxic metals co-exposure at levels relevant to the general population [e.g., median blood Pb was an order of magnitude lower than the reference value of 5 µg/dL (CDC 2020)] on anthropometric indicators measured in a relatively sizeable sample of US children selected from two NHANES cycles. Our analysis takes advantage of the recently developed nonparametric statistical approach BKMR, which is designed to flexibly assess dose–response, interactions, and mixture effects. BKMR may be poorly suited for very high-dimensional data; however, we only evaluated the exposure to a mixture of five metals (Lazarevic et al. 2019). Furthermore, our study explored potential variations by sex given differences in patterns of exposure, gastrointestinal absorption, and/or metabolism (Eriksson et al. 2010; Llop et al. 2013).

In conclusion, our findings suggest that simultaneous exposure to metals may influence children's body size and growth measurements, and that the effects of metals may differ depending on the exposure levels of other components in the metal mixture. Particularly, our main findings, based on a population of US school-age children, emphasize that further efforts are necessary to reduce Pb exposure from



drinking water and other potential sources while maintaining healthy levels of essential nutrients such as Mn to foster growth during the vulnerable period of childhood.

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## Compliance with Ethical Standards

**Conflict of interest** The authors declare they have no known competing financial interests of personal relationships that could have appeared to influence the work reported in this paper.

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